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Pesticides Ecotoxicological Risk Assessment for Surface Waters in the Cotton Growing Area Around the Bala's Hippopotamus Pond Biosphere Using PIRI Method

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Abstract

Pesticides residues are frequently found in the environment far from the original point of their application. Besides the desired effects of pest

control, non-target organisms, soil and water are contaminated by the pesticides. This paper presents results on the impact of these xenobiotics used in cotton cultivation on River "Wolo" environment in Burkina Faso by using the Pesticide Impact Rating Index (PIRI) software package. The assessment was based on the assumption of three scenarios taking into account the organic matter content of the soil and the presence of a buffer zone. Pesticides properties and use data, and data on the physical environment, were also used. Considering the worst case (scenario 2), diuron, haloxyfop-R-methyl, glyphosate and nicosulfuron were the most mobile. Diuron was classified as the most toxic pesticide to *Scenedesmus quadricauda*. Toxicity to *Daphnia magna* was extremely high with chlorpyrifos ethyl, very high with beta-cyfluthrin, deltamethrin, lambda-cyhalothrin and high with flubendiamide. For *Oncorhynchus mykiss*, it was beta-cyfluthrin, deltamethrin and lambda-cyhalothrin that caused a very high risk and chlorpyrifos ethyl and indoxacarb a high risk. For all pesticides, the risks are reduced overall depending on the width of the buffer zone and the organic matter content of the soil. The use of a pesticide in a given location must take into account its ecotoxicological impact on the surrounding ecosystem. Tools such as PIRI, could be used for the selection of pesticides to be used. Also, environmental parameters such as buffer zone and organic matter content should be used by farmers to limit the mobility of pesticides to water.

Keywords: Risks, pesticides, ecotoxicity, PIRI, Bala

Introduction

Conventional cotton production involves the massive use of different types of pesticides to control undesirable plants (herbicides), fungi (fungicides) and insect pests (insecticides) (Bayili et al., 2019, Gouda et al., 2018). The use of pesticides in cotton cultivation reduces crop losses (Miranda et al., 2013), which can be more than 50% (Traoré, 2008; Moussa, 2003; Silvie and Gozé, (1991) in the absence of phytosanitary treatments.

Although very useful, the abundant and abusive use of pesticides can lead to risks to human health (Shokrzadeh and Saravi, 2011; Fournier and Bonderef, 1983) and the environment (Calvet et al., 2005; Padovani et al., 2004). The risks to the environment concern not only contamination of air, soil and water but also toxicity to the non-target organisms. This toxicity is linked to contaminants that reach the living environment of living organisms through fairly complex processes. For pesticides, the process of water contamination seen by Morissette and Martel (2014) highlights a phenomenon of diffuse origin (erosion, runoff, leaching, preferential flow and drift) due to pesticide applications. Pesticides such as herbicides are applied directly to the soil, which increases the risk of leaching or erosion of pesticides into

groundwater or surface water (Taghavi, 2010; Aubertot et al., 2005). Under conditions of poor practice, poor management of rinse water could be associated, which can also contribute to contamination (Morissette and Martel, 2014). According to Mamy et al (2008), this contamination is influenced by:

- agro-pedoclimatic components such as soil texture and organic matter, rainfall, topography;
- technological components such as tillage, rotations, doses and frequency of pesticide applications; and
- the physicochemical properties of the pesticides used.

Knowledge of the environmental impact of pesticide use is based on the use of several methods. For example, direct and in situ measurements provide accurate data, but are generally expensive and difficult to implement and multiply (Bockstaller and Girardin, 2003; Mamy et al., 2008). Risk assessment methods, in this case indicators and models, help farmers and decision-makers to compare the environmental impact of different pesticides and to design effective control practices with minimal environmental impacts (Samuel et al., 2012; Muhammetoglu et al., 2010; Aravinna et al., 2005). One such method is the Pesticide Impact Rating Index (PIRI) which was developed by CSIRO (Kookana et al., 2005) to assess the potential impact of pesticides on water quality. It has already been used in Burkina Faso to assess the risk of surface and groundwater contamination by pesticides used in industrial sugar cane cultivation (Ouedraogo et al., 2012; Toe et al., 2012). However, a similar study has not been carried out on cotton crops to estimate the ecotoxicological potential of pesticides on specific aquatic ecosystems in adjacent environments. In cotton cultivation, the risks would be increased due to poor pesticide use practices (Bayili et al., 2019; Son et al., 2017; Tarnagda et al., 2017). The use of banned pesticides, overdosing, failure to maintain safe distances between fields and water points, lack of buffer zones, inappropriate mixing of pesticides, etc. are common practices among farmers.

The aim of this study was to assess the ecotoxicological impact of pesticides used for cotton growing on the aquatic systems of the river "Wolo" around the Bala's hippopotamus pond biosphere by using PIRI. The ecotoxicological potential was estimated for three aquatic organisms, the main links in the surface water trophic chain.

Methodology

Description of the study site

The study took place around the Bala's hippopotamus pond biosphere reserve, about 50 km north of the city of Bobo-Dioulasso in Burkina Faso (**figure 1**). This locality is part of the western zone of the "*Société Burkinabé des Fibres et Textiles* (SOFITEX)", the country's main conventional cotton

producer. This area is characterised by a Sudanian climate and an average rainfall of 986.7 measured during the study period. The minimum, maximum and average temperatures were 25.88°C, 28.01°C and 26.916°C respectively. In general, the soils are mainly tropical ferruginous soils with medium and shallow leaching, staining and indurated concretions, and hydromorphic soils with little humus and a pseudogley surface (BUNASOL, 2002). The hydrographic network of the reserve is made up of the Mouhoun River to the northwest and its tributary the "Wolo" to the southwest, which collects runoff from the cotton growing area towards the pond (Belem, 2008). The waters of the "Wolo" are the passive ones likely to be contaminated mainly by runoff, given the position and distance of this river from the cotton fields.

For the assessment, we monitored farming practices in four cotton fields adjacent to the river, from July to October during the 2018-2019 agricultural season. All the pesticide formulations used by these producers have been identified and characterized (table I). Site and stream characteristic data were collected (table II). The average soil organic matter content (1.4%), the erosion rate (0.4 t/ha/year) and the soil type were obtained from '*Bureau National des Sols du Burkina Faso*' (BUNASOL) databases (BUNASOL, 2002).

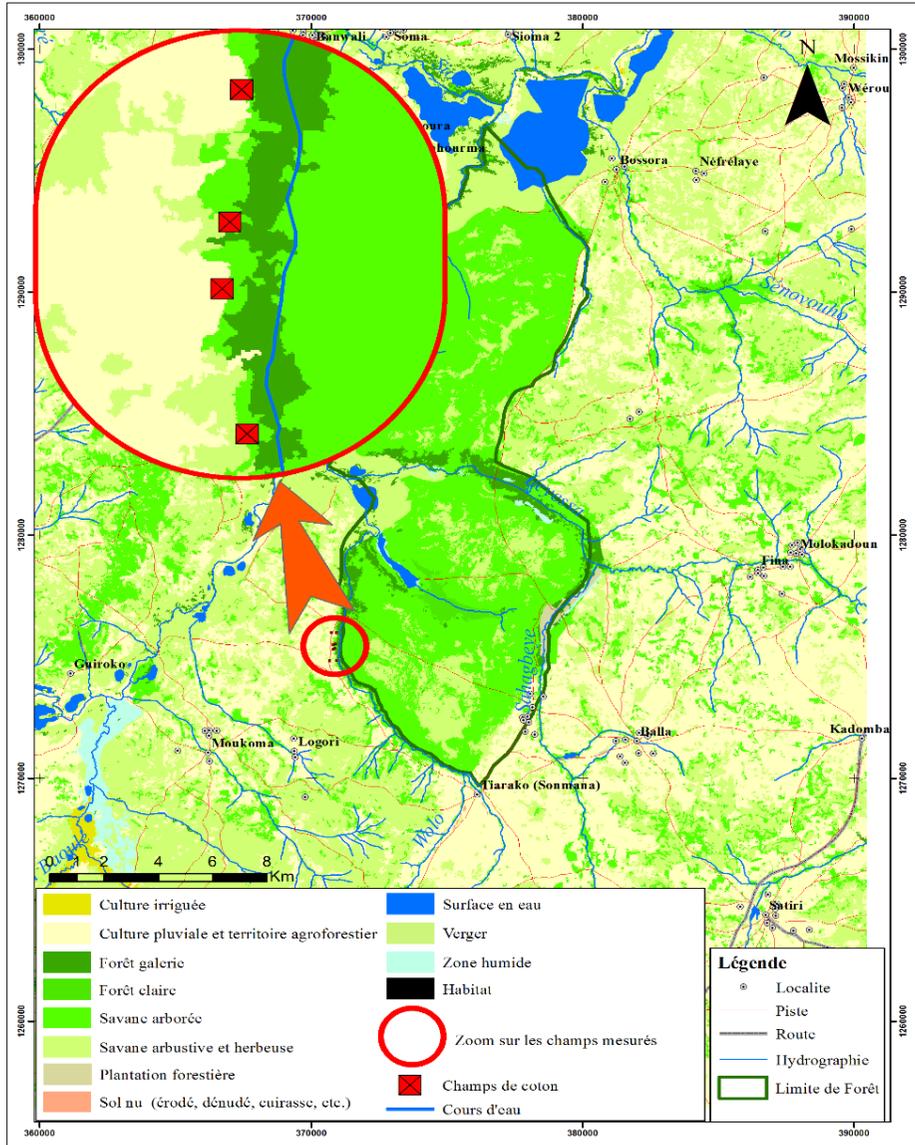


Figure 1: Location map of the study site, realized by Kinda (2019)

Pesticide Impact Rating Index (PIRI)

The Pesticide Impact Rating Index (PIRI) model was developed in Australia by the Commonwealth Scientific and Industrial Research Organization (CSIRO) to assess the potential impact of pesticides on water quality. It predicts the potential for chemical pesticides to move away from their place of application to pollute adjacent watercourses (CSIRO, 2001). The mobility of the pesticide and its effect on organisms is calculated by taking into account its toxicity, chemical properties, application rate and frequency.

Local soil and climate conditions are also taken into account (Kookana et al., 2005). The main transport routes for pesticides to surface water bodies are runoff, soil colloid erosion and spray drift. The pesticide concentration in the receiving water (CSW) is calculated from the pesticide load moving into the surface water adjacent to the treated area according to the following formula:

$CSW = L \times T \times WI / H$: with: **CSW** (in kg/m^3) the predicted concentration; **H** (m) the depth of the surface water body; **L**, the load of the pesticide applied to the soil; **T**, the overall surface transport coefficient for each pesticide; and **WI**, the water index defined as an approximate ratio of the length of the river bank adjacent to the perimeter of the area to be treated.

Impact assessment based on toxicity for an organism is made by considering a series of aquatic organisms represented by different trophic levels. These are the LC50 for rainbow trout used as a measure of toxicity to fish, the LC50 for *Daphnia sp.* and the EC50 for algae (Kookana et correll, 2008; Kookana et al, 2005). The negative impact of the amount of pesticide moving away from the spray site is governed by the concentration of the pesticide in the receiving environment (Cws) relative to the concentration that is toxic to the exposed organism. The risk index for surface water is calculated by the model from the predicted concentration (CSW) and the toxicity value (EC50 or LC50).

Risk index for surface water = Csw / EC50 or LC50

Each pesticide is classified by PIRI in one of the following risk categories: "very low", "low", "moderate", "high", "very high", and "extremely high" according to its ecotoxicological potential, more than another pesticide that is also subject to the same site conditions (Kookana and Correll, 2008).

For the present study, the reference species "*daphnia magna*" for crustaceans, "*Oncorhynchus mykiss*" for fish and "*Scenedesmus quadricauda*" for algae were chosen according to their ecological importance and the availability of ecotoxicological data. For the active ingredients (table I), all KOC, DT50, crustacean LC50, and most fish LC50 and algae EC50 values have been provided by the Pesticide Properties Database (Footprint PPBD, 2020).

Table I: Technological, physico-chemical and ecotoxicological characteristics of pesticides

Pesticide formulation	Active ingredient (a.i.)	Concentration a.i. (g/L or g/kg)	Application rate (kg or L/ha)	Koc (mL/g)	LC50 (mg/L)		EC50 (mg/L) <i>Scenedesmus quadricauda</i>	DT50 (days)
					<i>Oncorhynchus mykiss</i>	<i>daphnia magna</i>		
ACERO 84 EC*	Isoclast /sulfoxaflor	48	0.5	40.8	> 101	> 399	> 101	2.2
	Lambda-cyhalothrin	36	0.5	283707	0.00021	0.00036	> 0.3	175
ADWUMA WURA 480 SL**	Glyphosate	480	2	1424	38.0	40	> 72.9	15.0
AVAUNT 150 EC*	Indoxacarb	150	0.17	17.6	> 0.17	0.17	>108 ¹	113.2
DIURALM 80 WG**	Diuron	800	1	813	4.9 ²	5.7	0.0027	75.5
GALLANT SUPER**	Haloxypop-R-methyl	104	0.9	66	0.46 ³	12.3	> 94.9 ³	0.5
GLYPHADER 360 SL**	Glyphosate	360	2	1424	38.0	40	4.4	15.0
GRAMOSHARP SUPER**	Paraquat dichloride	200	0.5	1000000	19	4.4	0.04 ⁴	365
INDOXAN*	Indoxacarb	50	0.5	17.6	> 0.17	0.17	0.079	113.2
KILLER**	Glyphosate	360	4	1424	38.0	40	> 72.9	15.0
NICOMAIS**	Nicosulfuron	40	1.5	30	65.7	90.0	3.7 ⁵	26
POWER 80 WG**	Diuron	800	4	813	4.9	5.7	0.0027	75.5
PYRINEX QUICK 424EC*	Deltametrin	24	0.5	10240000	0.00026	0.00056	2.560 ⁶	13
	Chlorpyrifos ethyl	400		5509	0.025	0.00010	0.660 ⁷	386
THUNDER*	Imidacloprid	100	0.2	255	> 83	85	> 10 ⁸	191
	Beta-cyfluthrin	45		64300	0.000068	0.00029	>0.01 ⁹	13
TIHAN 175 O-TEQ*	Spirotetramat	75	0.22	289	2.54 ¹⁰	> 42.7	0.36 ¹¹	0.19
	Flubendiamide	100		2197	0.06	0.06	>0.069	500

*insecticide ; **herbicide ; ¹ECHA, 2020 ; ²Fojut *et al.*, 2011 ; ³FAO, 2012 ; ⁴Sáenz *et al.*, 1993 ; ⁵PubChem, 2020a ; ⁶PubChem, 2020b ; ⁷NRA, 2020 ; ⁸CCME, 2007 ; ⁹FAO, 2016 ; ¹⁰Agbohessi *et al.*, 2013 ; ¹¹ANLA, 2018.

Scenarios

- Scenario 1 (actual case observed) where the soil organic matter content is 1.4%, with a buffer zone;
- Scenario 2 where the soil organic matter content was maintained at 1.4% and a 5 m non-vegetated zone was assumed;
- Scenario 3 includes a soil organic matter content of 2%, a distance between the field and the water course of 100 m (Agence de l'Eau du Nakanbé (AEN), 2015) and a 5 m grassed buffer zone.

Table II: Characteristic data of the site and the river "Wolo".

Measured parameters	Values
Distance from the edge of the field to the water surface (m)	75.25
Average river diameter (m)	5.625
Average depth of river (m)	1.625
Erosion rate (t/ha/year)	0.4
Slope	1.89
Type of soil	Clay-silt
Organic matter content (%)	1.4
Total rainfall during the period (June to October) (mm)	986.7
Average minimum temperature during the period (°C)	21.82
Average maximum temperature during the period (°C)	33.05
Minimum number of days between the application of pesticides and the first rainfall	2

Results and discussion

Potential for mobility

Mobility (table III) is high for Diuron, Haloxyfop-R-methyl, glyphosate and nicosulfuron in scenario 2. Adsorption coefficients (Koc) indicate that glyphosate and diuron are slightly mobile while nicosulfuron and Haloxyfop-R-methyl are mobile. Their mobility is very low in Scenario 1, where the width of the buffer zone (75.25 m) was the determining factor for this decrease. Thanks to its phytoremediation capacity, the vegetated strip of the buffer zone has a great influence on reducing the risk of water pollution by chemical pesticides (Trainer and Volker, 2008). This parameter acts as a brake on erosion, which is one of the factors contributing to the pollution of surface waters by pesticides. In scenario 3, where only diuron has a high mobility, the overall risk reduction is influenced by the increase in soil organic matter content (2%). Organic matter increases the activity of microorganisms that have the power to degrade organic pesticides in general. It adsorbs pesticides, increases their residence time in soils and promotes their degradation by microorganisms (Savadogo et al., 2006). It also accelerates the degradation of pesticides in the soil (Savadogo et al., 2008). Since microorganisms are the former agents for the degradation of organic contaminants in soil, the

application of organic matter, which increases microbial density and also provides nutrients and readily degradable organic matter can be considered useful to accelerate the contaminant degradation (Takeshita *et al.*, 2019; Gómez *et al.*, 2014; Masciandaro *et al.*, 2013). Moreover, the organic matter addition, by means of the increase of cation exchange capacity, soil porosity and water-holding capacity, enhances the soil health and provides a medium satisfactory for microorganism activity. The buffer zone and organic matter could offer solutions within the framework of a risk management plan for pesticide mobility (Calvet *et al.*, 2005; Ouedraogo *et al.*, 2012).

Table III: Classification of pesticide mobility

Pesticide	Scenario 1	Scenario 2	Scenario 3
Diuron	Very low	High	High
Haloxypop-R-methyl	Very low	High	Medium
Glyphosate	Very low	High	Medium
Nicosulfuron	Very low	High	Medium
Sulfoxaflor	Very low	Medium	Medium
Imidacloprid	Very low	Medium	Medium
Chlorpyrifos ethyl	Very low	Medium	Medium
Paraquat dichloride	Very low	Medium	Low
Flubendiamide	Very low	Low	Low
Indoxacarb	Very low	Low	Very low
Lamda-cyhalothrin	Very low	Very low	Very low
Deltamethrin	Very low	Very low	Very low
Beta-cyfluthrin	Very low	Very low	Very low
Spirotetramat	Very low	Very low	Very low

Impact on Scenedesmus quadricauda

The classification of the impact of pesticides on algae is given in table IV. Diuron causes an extremely high risk in scenarios 2 and 3 and a very high risk in scenario 1. The other pesticides induce a toxicity ranging from very low to medium according to the different scenarios. The high estimated mobility and intrinsic toxicity of diuron towards algae are at the origin of the risk of this herbicide. This level of toxicity of this pesticide towards algae could threaten the health and productivity of the aquatic ecosystem, given the importance of phytoplankton in the trophic chain. The data from Scenario 1 show that risk reduction was influenced by the width of the buffer zone, when the data from this scenario are compared overall with those from the other scenarios. Indeed, except for diuron, the risks are low to very low for all pesticides.

Table IV: Classification of pesticide toxicity to *Scenedesmus quadricauda*

Pesticide	Scenario 1	Scenario 2	Scenario 3
Beta-cyfluthrin	Very low	Medium	Medium
Chlorpyrifos ethyl	Very low	Medium	Medium
Deltamethrin	Very low	Very low	Very low
Diuron	Very high	Exc. high	Exc. high
Flubendiamide	Very low	Medium	Medium
Glyphosate	Very low	Very low	Very low
Haloxypop-R-methyl	Very low	Very low	Very low
Imidacloprid	Very low	Low	Low
Indoxacarb	Very low	Very low	Very low
Lamda-cyhalothrin	Very low	Very low	Very low
Nicosulfuron	Very low	Low	Low
Paraquat dichloride	Low	Medium	Medium
Spirotetramat	Very low	Very low	Very low
Sulfoxaflor	Very low	Very low	Very low

Impact on Daphnia magna

Table V presents the results of the impact on *Daphnia*. In scenario 2, toxicity is extremely high with chlorpyrifos ethyl, very high with beta-cyfluthrin, deltamethrin, lamda-cyhalothrin and high with flubendiamide. All these pesticides are used for their insecticidal effects. The intrinsic toxicity of these pesticides is high towards *Daphnia* (table I). Despite mitigation measures (in scenarios 1 and 3), the overall toxic impact of these insecticides remains a cause for concern. Indeed, chlorpyrifos-ethyl and lambda-cyhalothrin have very high and high toxicity respectively in scenario 1. The use of these potentially toxic pesticides around areas of particular ecological importance, such as the hippopotamus pond biosphere reserve (RBMH), must be strongly controlled in order to preserve the aquatic ecosystem.

Table V: Classification of pesticide toxicity to *Daphnia magna*.

Pesticide	Scenario 1	Scenario 2	Scenario 3
Beta-cyfluthrin	Medium	Very high	Very high
Chlorpyrifos ethyl	Very high	Exc. high	Exc. high
Deltamethrin	Medium	Very high	High
Diuron	Very low	Medium	Medium
Flubendiamide	Very low	High	Medium
Glyphosate	Very low	Low	Very low
Haloxypop-R-methyl	Very low	Medium	Low
Imidacloprid	Very low	Very low	Very low
Indoxacarb	Very low	Low	Low
Lamda-cyhalothrin	High	Very high	Very high
Nicosulfuron	Very low	Very low	Very low
Paraquat dichloride	Very low	Very low	Very low

Spirotetramat	Very low	Very low	Very low
Sulfoxaflor	Very low	Very low	Very low

Impact on *Oncorhynchus mykiss*

The results of the impact assessment on this fish are in table VI. Considering scenario 2, beta-cyfluthrin, deltamethrin and lamda-cyhalothrin cause a very high risk while chlorpyrifos ethyl and indoxacarb cause a high risk. For all these pesticides, the level of risk remains constant in scenario 3. Nevertheless, the general trend is that the level of risk for most other pesticides is decreasing. The risk ranking for these latter pesticides is in the lower order in this scenario 3, ranging from moderate to very low. The increase in organic matter content and the presence of the buffer zone have caused this decrease. Although the width of the buffer zone in scenario 1 caused a reduction in risk for all pesticides, this is not the case for beta-cyfluthrin and lamda-cyhalothrin for which the risk is high. The classification of these pesticides is greatly influenced by their higher intrinsic toxicity than all other pesticides (table I).

Table VI: Classification of pesticide toxicity to *Oncorhynchus mykiss*

Pesticide	Scenario 1	Scenario 2	Scenario 3
Beta-cyfluthrine	High	Very high	Very high
Chlorpyrifos ethyl	Medium	High	High
Deltamethrin	Medium	Very high	Very high
Diuron	Very low	Medium	Medium
Flubendiamide	Very low	Medium	Medium
Glyphosate	Very low	Very low	Very low
Haloxypop-R-methyl	Very low	Low	Very low
Imidacloprid	Very low	Very low	Very low
Indoxacarb	Very low	High	High
Lamda-cyhalothrin	High	Very high	Very high
Nicosulfuron	Very low	Very low	Very low
Paraquat dichloride	Very low	Very low	Very low
Spirotetramat	Very low	Very low	Very low
Sulfoxaflor	Very low	Very low	Very low

Conclusion

The level of ecotoxicological risk of pesticides on the waters of the "Wolo" River was estimated using the PIRI model. Considering the worst case (scenario 2), diuron, haloxypop-R-methyl, glyphosate and nicosulfuron were the most mobile. Diuron was classified as the most toxic pesticide to *Scenedesmus quadricauda*. Toxicity to *Daphnia magna* was extremely high with chlorpyrifos ethyl, very high with beta-cyfluthrin, deltamethrin, lamda-cyhalothrin and high with flubendiamide. In *Oncorhynchus mykiss*, it was beta-cyfluthrin, deltamethrin and lamda-cyhalothrin that caused a very high risk and chlorpyrifos ethyl and indoxacarb a high risk. Pesticides have shown a

differential level of risk depending on their mobility and their toxic properties on organisms. Diuron for algae, chlorpyrifos ethyl and lambda-cyhalothrin for *Daphnia* and beta-cyfluthrin and lambda-cyhalothrin for fish induced at least high toxicity in all scenarios. In general, the association of the buffer zone with the increase in soil organic matter content led to a decrease in the level of risk. The use of a pesticide in a given location must take into account its ecotoxicological impact on the surrounding ecosystem. Tools such as PIRI, could be used for the selection of pesticides to be used. Also, environmental parameters such as buffer zone and organic matter content should be used by farmers to limit the mobility of pesticides to water.

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